

Review

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Revisiting the relationships between human well-being and ecosystems in dynamic social-ecological systems: Implications for stewardship and development

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Non-technical summary. We argue that the ways in which we as humans derive well-being from nature – for example by harvesting firewood, selling fish or enjoying natural beauty – feed back into how we behave towards the environment. This feedback is mediated by institutions (rules, regulations) and by individual capacities to act. Understanding these relationships can guide better interventions for sustainably improving well-being and alleviating poverty. However, more attention needs to be paid to how experience-related benefits from nature influence attitudes and actions towards the environment, and how these relationships can be reflected in more environmentally sustainable development projects.

Technical summary. In the broad literatures that address the linked challenge of maintaining ecosystem integrity while addressing poverty and inequality, there is still a need to investigate how linkages and feedbacks between ecosystem services and well-being can be taken into account to ensure environmental sustainability and improved livelihoods. We present a conceptual model towards a dynamic and reciprocal understanding of the feedbacks between human well-being and ecosystems. The conceptual model highlights three mechanisms through which people derive benefits from ecosystems (use, money and experience), and illustrates how these benefits can affect values, attitudes and actions towards ecosystems. Institutions and agency determine access to and distribution of benefits and costs, and also present barriers or enabling factors for individual or collective action. The conceptual model synthesises insights from existing but mostly separate bodies of literature on well-being and the benefits humans derive from ecosystems, and reveals gaps and areas for future research. Two case studies illustrate how recognizing the full feedback loop between how ecosystems support human well-being and how people behave towards those ecosystems, as well as intervention points within the loop, can guide better action for sustainable poverty alleviation and stewardship of the biosphere.

1. Introduction

Ecosystems play a critical role in underpinning the well-being of humanity (Folke *et al.*, 2016; MA, 2005). However, human actions are rapidly modifying the structure and function of the Earth's ecosystems and reducing their potential to support human well-being (Steffen *et al.*, 2015). Rural communities in developing countries are particularly vulnerable to ecological change due to their direct dependence on ecosystems and their services (Suich *et al.*, 2015). A major challenge, today and into the future, is to maintain or enhance beneficial material and non-material contributions of nature to a good quality of life for all people. Indeed, the UN Agenda 2030 and the Sustainable Development Goals (Griggs *et al.*, 2013; Sachs, 2012; UN, 2015) highlight the linked challenge of maintaining ecosystem integrity while addressing poverty and inequality. This challenge requires institutions, behaviours and governance systems that support both benefits from ecosystems to people, and the stewardship of those ecosystems (Guerry *et al.*, 2015).

Since the Millennium Ecosystem Assessment, a proliferation of studies has contributed to the documentation, classification and understanding of how a wide range of 'ecosystem services' contribute to human well-being (Cruz-García *et al.*, 2017; Díaz *et al.*, 2015; Rounsevell *et al.*,

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2010; Suich *et al.*, 2015; TEEB, 2010). In particular, there has been progress in the literature on the multidimensional nature of human well-being and the different ways in which ecosystem services contribute to enhancing different objective and subjective components of human well-being (Breslow *et al.*, 2016; Pascual *et al.*, 2017; Russell *et al.*, 2013). Engagement with different cultural systems and worldviews has also questioned and broadened conceptualizations and measures of well-being and development (Díaz *et al.*, 2018; Diener & Suh, 2000; Pascual *et al.*, 2017; Sterling *et al.*, 2017a). Indeed, emerging emphasis on values and culture in the Intergovernmental Science-Policy Platform for Biodiversity and Ecosystem Services (IPBES) illustrates the increased recognition of the importance of human–nature relations in maintaining well-being and protecting biodiversity (Díaz *et al.*, 2015, 2018; Pascual *et al.*, 2017). Recent advances have explored cultural aspects of these benefits, going beyond aesthetic and recreational values to also recognize place attachment, sense of belonging, rootedness and spiritual connections such as links between ancestors, spirits and gods to certain places or natural features (Chan *et al.*, 2016; Cocks *et al.*, 2016, 2012; Pascua *et al.*, 2017; Verschuuren *et al.*, 2014). Others have explored how various social structures and mechanisms of access determine who benefits from ecosystems (Berbés-Blázquez *et al.*, 2017; Berbés-Blázquez *et al.*, 2016; Daw *et al.*, 2011; Pascual *et al.*, 2014).

At the global level, few studies have investigated how linkages and feedbacks between ecosystem services and well-being determine environmental sustainability and improved livelihoods, with a particular lack of research in the global South (Cruz-García *et al.*, 2017). Furthermore, some human well-being frameworks, such as the international quality of life surveys, do not consider the natural environment as a determinant or constituent of well-being at all (Schleicher *et al.*, 2018). This is an especially concerning research gap for developing countries, challenged with multidimensional poverty and high rates of ecosystem loss (Cruz-García *et al.*, 2017; Suich *et al.*, 2015). In efforts to reduce poverty through economic growth, it is critical that we do not sever other, more intangible connections to nature that matter for long term sustainability of ecosystems on which people depend (Haider *et al.*, 2018; Schleicher *et al.*, 2018).

Such a fragmented, disconnected understanding of the complex dynamics between well-being and ecosystems hinders the development of effective stewardship and governance of ecosystems (Folke *et al.*, 2016; Summers *et al.*, 2012). A holistic consideration of the range of ways in which people derive well-being from ecosystems, and how this feeds back to the ways in which people value ecosystems and engage in stewardship and governance of ecosystems, could reveal opportunities for better governance or poverty alleviation and development interventions.

In this paper, we draw on different sets of literature to present a synthesized conceptual model of the structure and operation of this hypothesized ‘feedback loop’. We illustrate aspects of this feedback loop and the potential implications for interventions through two case studies, one thematic case on rangeland management in Southern Africa, and a regional example from West Kalimantan, Indonesia. A social-ecological systems (SES) framing draws direct attention to the need to understand how human–nature relationships can influence stewardship and maintenance of benefits (Folke *et al.*, 2016), and how institutions mediate these linkages in different ways. These insights are informed and inspired by our experiences of place-based research of human–nature relationships in case studies of resource-dependent communities, within the community of practice of the Programme for

Ecosystem Change and Society (PECS)ⁱ. While we draw upon the concept of ecosystem services in this paper, we acknowledge that it has become a dominating approach to conceptualize human–nature interactions in the science–policy interface, and has been criticized as masking other ways of study, including the full range of research on human–nature interactions in the social sciences and the humanities (Díaz *et al.*, 2018; Stenseke, 2016). We take a holistic view of ecosystem services and consider both tangible and intangible benefits that can be obtained from nature. We further broaden the discussion by incorporating an understanding of consequent behaviour and actions of resource users as a function of multi-scalar institutional constraints, political capacity, power relations, and cultural and economic values. These aspects are well developed in the existing field of political ecology, and furthermore echo related insights in the disciplines of anthropology, geography and sociology. Finally, we conclude by reflecting on key emerging research questions, to inform development and poverty alleviation interventions that strengthen well-being and connections to nature, and improve sustainability.

2. Conceptual model: The feedback between ecosystems and well-being

We present a conceptual model (Figure 1; see also Box 1 for use of terms) that brings together different strands of literature to highlight a potential feedback loop connecting mechanisms by which people benefit from nature to their attitudes and behaviours towards local ecosystems, through their material and perceived well-being. This feedback loop could operate at both an individual and collective level. Most importantly, it is mediated through institutions that might structure, enable or constrain individual agency, or reflect the aggregate behaviours of many, or few powerful individuals. Such reinforcing feedbacks have the potential to perpetuate rapid changes in a system (Chapin *et al.*, 2009). For example, a reduction in well-being benefits due to a decline in ecosystem quality may enhance exploitation and diminish stewardship, which can lead to increased ecosystem degradation and a further decline in benefits (a poverty–environment trap as described by, e.g., Barbier 2010). We also know that under certain conditions, people can and do act as stewards of their dynamic resources and ecosystems through active use and management (Berkes *et al.*, 2003; Comberti *et al.*, 2015; Ostrom, 1990; also see Box 1). However, a range of social and institutional factors influencing the generation and distribution of ecosystem services on the one hand, and the potential for environmental stewardship on the other, affect how this feedback operates. We draw on existing literature and theory to point to the key factors that could affect such a feedback mechanism and initiate a discussion of the conditions (e.g. historical context, and power relations that shape institutions and agency) which may affect or reinforce patterns of behaviour that result in degradation of ecosystems or in governance systems that steward ecosystems. In the following, we elaborate further on the different parts of the conceptual model outlined in Figure 1.

2.1. The benefit basket and mediating factors

Our model illustrates three mechanisms (see point 2, Figure 1) by which benefits from ecosystems are derived (based on Chaigneau *et al.*, unpublished observations). Ecosystems can be used directly (e.g. fish may be caught for household consumption); they can provide a monetary income (e.g. caught fish can be sold at market); or they may constitute an experience (e.g. the act of fishing

Box 1. Glossary.

Agency: the capacity of individuals and groups to act independently to make their own free choices (Brown & Westaway, 2011). Agency can be determined or limited by structure, which includes factors such as gender, ethnicity, customs, class etc. Within a social-ecological systems approach, agency should be considered in a dynamic interplay with structure, where “emergent political processes reflect both the agency of current actors and the influence of historically embedded structures, practices and legacies” (Ezzine-De-Blas et al., 2015, p. 73).

Attitudes: an evaluation of (including beliefs, feelings and behavioural tendencies towards) a mental or physical object (i.e. a person, place, thing, issue, event or symbol) (Eagly & Chaiken, 1993; Hogg & Vaughan, 2005). The relationship between attitudes and behaviour is often described by the theory of reasoned action, which states that attitudes influence behavioural intentions, which in turn influence actions (Ajzen & Fishbein, 1980; Casaló & Escario, 2018). More broadly, the cognitive hierarchy model links values via attitudes to behaviour. In this model, attitudes mediate the influence of values on specific behaviours (Homer & Kahle, 1988). In general, attitudes refer to specific entities (i.e. person, thing or issue), whereas values have no object of reference (Rokeach, 1973).

Benefit basket: a set or bundle of benefits that are derived from ecosystems through the different mechanisms of use, money or experience (after Chaigneau et al., unpublished observations).

Ecosystem services: All of the contributions, both positive and negative, of living nature (including its diversity of organisms, ecosystems, and their associated ecological and evolutionary processes) to people’s quality of life (MA, 2005). Ecosystem services are co-produced by the interactions between nature and people (Palomo et al., 2016; Reyers et al., 2013).

Human well-being: a multidimensional concept, encompassing a number of factors that contribute towards a person’s or group’s condition and sense of well-being. These factors include the basic material for a good life (e.g. food, income, shelter), physical and mental health, good social relations (e.g. positive relationships with family and friends), security (e.g. personal safety) and freedom of choice and action (e.g. personal fulfilment and freedom of expression) (MA, 2005). Certain dimensions of well-being may be assessed using objective measures such as income. However, well-being is also a highly subjective concept that depends on a person’s life satisfaction and happiness (Diener, 1984).

Institutions: The humanly devised constraints that structure political, economic and social interaction. They consist of both informal constraints (sanctions, taboos, customs, traditions and codes of conduct) and formal rules (constitutions, laws, property rights) (North, 1991).

Stewardship: the responsible use and management of ecosystems and their natural resources (Enqvist, 2017). We use stewardship in a broad sense that can describe actions as well as systems of management or governance, and include a range of actors at local level or higher up as embedded in decision making (in contrast to, e.g., ‘local environmental stewardship actions’ as defined by Bennett and colleagues (2018)).

Values: Schwartz (1994, p. 20) building on Rokeach’s work on values (Rokeach, 1973) defines a value as a “belief pertaining to desirable end states or modes of conduct that transcends specific situations, guides selection or evaluation of behaviour, people and events.” Values are more central and deeply held than attitudes and underpin decisions and behaviours (Jones et al., 2016; Satterfield, 2001).

may contribute to a sense of joy, identity, freedom, autonomy and being respected by others). Of the three mechanisms, the importance of monetary and direct use benefits (mostly of provisioning services) are most widely recognized in ecosystem services literature from the global South (Suich et al., 2015) and in development interventions. In contrast, non-material benefits derived through experience have traditionally received relatively less attention in the ecosystem services/human well-being literature, as non-material aspects do not easily lend themselves to quantitative assessments (Chan et al., 2012; Russell et al., 2013).

The three benefit mechanisms are not mutually exclusive, and can be part of the same ‘benefit basket’. In the Eastern Cape province of South Africa, for example, collecting firewood for cooking not only provides materials that contribute to making meals, but also fosters social cohesion among the women who share the experience of collecting the wood, contributes to self-respect and a sense of identity as hard-working women, and provides them with an opportunity to escape the troubles of home and connect with ancestral spirits in the forest (Cocks et al., 2012). Different benefit mechanisms may therefore contribute directly to different human well-being domains, for example, physical security or health (Doyal & Gough, 1991; Chaigneau et al., unpublished observations). However, there may be trade-offs between the benefit mechanisms – benefitting from one mechanism may limit or reduce the chances of benefiting from another mechanism from the same ecosystem, potentially creating

inequities between different beneficiary groups. For example, setting aside forest areas for ecotourism creates opportunities to earn income for some people in a community, but reduces the access for other groups to harvest and use resources and derive the associated experiential benefit. The benefits of particular ecosystem services, and the costs associated with producing them (e.g. loss of alternative ecosystem services), are often distributed unequally amongst people (Daw et al., 2011).

The distribution of benefits and costs is determined by social structures and mechanisms of access that operate at the individual and societal levels (see red circles in Figure 1). At an individual level, a person’s agency (i.e. a person’s capacity to act) in accessing ecosystem services is determined by a range of means, relations and processes (Ribot & Peluso, 2003), including entitlements, technology, capital, markets, labour, knowledge, authority, social relationships and social identity (Leach et al., 1999; Palomo et al., 2016; Reyers et al., 2013). At a collective level, institutions such as policies, tenure systems, cultural norms, as well as the discourses of and the power relations within these institutions, determine who has the potential to access ecosystem services (Berbés-Blázquez et al., 2016; Escobar, 1996; Hajer, 1995). For example, cultural norms around gender often determine the ecosystem services potentially available to men or women and the means to access these services (Djoudi & Brockhaus, 2011). In East Africa, commercial mangrove pole cutting is done exclusively by men, and the poles are predominantly perceived as important

Fig. 1. Integrative conceptual model illustrating the dynamic interrelation between ecosystems (1), their benefits to people (2), well-being (3), and people's values and actions (4–5) that affect the condition of ecosystems. On the right hand side (2), three main mechanisms by which benefits from ecosystems are derived make up a 'benefit basket' and contribute to ecosystem-supported constituents of well-being. Three key components of the benefit basket are highlighted: direct use, exchange for money and experience. In reality they are not distinct but interacting and interdependent. Illustrated on the left hand side, is how the well-being derived from these benefits, and people's perception thereof, influence individual and collective attitudes and values towards nature (4), and ultimately the behaviours and actions (5) taken in relation to ecosystems to maintain or enhance benefit. The scattered indications of the benefit basket components illustrate that ecosystem-derived benefits become part of a broader mix of well-being factors that influence values and action. Agency and institutions (red circles) on the right side of the loop refer to factors that mediate the benefits from ecosystems, such as allocation of, and access to resources (which can be restricted by gender, class, race and historical inequities), as well as capacity and willingness to acquire the benefits. On the left hand side, agency and institutions relate to what actions are possible and for whom. Illustration by J. Lokrantz/Azote.



for monetary benefit. Women, on the other hand, collect firewood in the mangroves for home use and enjoy the relationships fostered by the experience of collecting together. Consequently, women predominantly perceive this service to be important as a direct use and as an experiential benefit. Institutions reflect historical and contemporary power relations, and also influence the relative power between different social groups, often resulting in inequity in the distribution of benefits and costs from the use of ecosystem services (Berbés-Blázquez et al., 2017). In South Africa, Kepe et al. (2015) found that initiation into manhood, which requires seclusion in undisturbed wooded areas, took precedence over women's access to firewood and led to conflict in urban areas where wooded areas were scarce. Institutions and agency therefore influence access to resources and potential benefit baskets at both the individual and collective level.

2.2. From benefits to values, attitudes and behaviours

The left-hand side of Figure 1 (points 3–5) illustrates how people's well-being, and more specifically their perception of nature's contribution to well-being, can impact their attitudes, values and ultimately behaviour towards the natural environment. In general, experiences of nature, and of its management and destruction, have been identified in the literature as one important predictor of individuals' 'pro-environmental behaviour' (Kollmuss & Agyeman, 2002). However, the effect is not linear and involves complex interactions with other factors and trade-offs with other values (Klößner, 2013).

For one, behaviour may be determined by both the subjective and objective aspects of people's well-being (Coulthard, 2012). Well-being is a broad and multidimensional concept that does not depend solely on benefits derived from ecosystems. Nevertheless, well-being is increasingly argued to be linked to nature and the benefits it provides and this is an area that needs further inquiry (Díaz et al., 2015, 2018; MA, 2005; Schleicher et al., 2018). In this conceptual model, we focus particularly on those aspects of well-being that are supported by ecosystems (Figure 1, point 3). The lower section of Figure 1 reflects evidence that the different ways through which ecosystems are perceived to support well-being can inform attitudes, values, beliefs and emotions that people individually or collectively hold with regard to nature (Gavin et al., 2015; Jones et al., 2016; Loomis & Paterson, 2014; Sterling et al., 2017a). Beyond absolute material or immaterial benefits, these attitudes are also affected by the perceived fairness of how benefits are shared. For example, jealousy over unequal benefits from Filipino marine protected areas was a key motivation behind poaching behaviour (Chaigneau, 2013).

Points 4 and 5 in Figure 1 indicate how these attitudes and values can, in turn, influence actions towards stewardship of ecosystems. Environmental psychology has shown that environmental attitudes and values are an important factor influencing intentions and thus behaviour (Klößner, 2013). However, it is well known that attitudes do not necessarily translate into actions, especially when it comes to pro-environmental behaviour (Dietz et al., 2005; Kollmuss & Agyeman, 2002; Stern et al., 1995; Tanner, 1999; Whitmarsh, 2009). In fact, individual and collective

Box 2. Managing the rangeland commons: Why such an elusive goal?

Sustainable management of communal rangelands in semi-arid regions presents an ongoing conundrum to policy makers (Vetter, 2013). Usually, a communal grazing area is evaluated as being overgrazed, degraded and unproductive, and interventions (historically top-down but increasingly participatory) are designed to improve rangeland condition and economic gain. Interventions typically take the form of improved grazing systems and improved livestock breeds coupled with economic incentives such as marketing of livestock. With rare exceptions, these fail to gain broad-based support and are usually short-lived. When assessing these approaches against our conceptual framework, several biases and gaps become apparent:

1. Interventions focus on economic productivity and environmental sustainability of rangeland use but usually fail to consider the full **benefits basket** (see point 2 in Figure 2). The multiple ways in which rangelands benefit different people through use and monetary exchange are becoming better understood (Hebinck & Shackleton, 2010; Shackleton et al., 2001; Shackleton et al., 2008), but are often not included in rangeland management interventions. The experiential, including cultural, dimensions of livestock ownership and other rangeland uses are poorly understood outside the anthropological literature (Ainslie, 2005).
2. Even though there is increasing awareness of **access issues and inequalities** (such as highly skewed livestock ownership, unequal power dynamics between men and women), these factors are seldom effectively considered and addressed (right-hand red circle, Figure 2). Interventions tend to target cattle owners, who are often a minority, that is, relatively wealthy, older men. The interests of women and other people who do not own livestock, but still derive benefits from the common resource, are usually under-represented.
3. Participatory assessments to elicit perceptions of environmental problems and development needs have become commonplace, and often reveal some recognition of land degradation. However, very little is known about the **attitudes and values** of livestock owners and other users of the rangeland commons (see point 4 in Figure 2). Consequently, in-depth and multifaceted understandings of people's feelings about, and attachment to, their natural environment are rare and seldom inform development interventions.
4. There is often a notable mismatch between land users' expressed sentiments about the natural environment (e.g. recognition that it has degraded over time) and their **behaviours/actions** (see point 5 in Figure 2). This commonly takes the form of non-compliance with grazing regulations, but also a lack of uptake of the economic opportunities meant to increase benefits such as selling livestock when market access is improved. The reasons for this are poorly understood. Lack of compliance with grazing regulations may be due to constraints such as water availability, or a lack of locally supported institutions and cooperation (see 4–5 in Figure 2). The seeming reluctance to sell livestock, particularly cattle, is also puzzling if the benefits of livestock are understood purely in economic terms.

There has been considerable research to understand the **structural barriers** to effective common property management and the constraints limiting the uptake of opportunities made available by development projects. For example, cattle are deeply embedded in economic and kinship networks, as well as having cultural significance, which means that they are not simply an economic good to be traded or disposed of (Ainslie, 2005). At this point, however, research and interventions tend to be informed by particular lines of evidence (e.g. rangeland ecology but not anthropology and political studies), and as a result target only selected parts of the feedback loop, that is, the monetary benefits and economic incentives for compliance with grazing regulations.

We propose that explicit and careful consideration of the full loop as proposed in our conceptual model would help address this problem and provide guidance for more appropriate and sustainable stewardship and development. For example, understanding the full range of benefits from rangelands and how different people derive them can inform interventions that increase the total benefit basket and benefit people more equitably. This, together with a better understanding of how people perceive and value these multiple benefits, and their relationship with the natural environment, could inform more appropriate strategies for involving people in better resource management.

behaviours are constrained (or enabled) by individual agency and the social institutions, rules and norms (see red circle in Figure 1) that govern behaviour around natural resource use and management (e.g. taboos concerning extractive resource use in certain forests in southern Madagascar (Tengö et al., 2007)). In understanding the interplay between institutional structures and agency, it is critical to recognize the constraints within which individuals and local actors can manoeuvre. Interventions (and values) are thus locally adapted, re-crafted and embedded in everyday practices (Van Hecken et al., 2015). Actions affecting ecosystems in any given setting respond not just to individual perceptions of the ecological dynamics and the benefit basket, but also to an institutional context shaped by particular types of knowledge, ways of doing, habits and norms that have been moulded historically by relations of power (Berbés-Blázquez et al., 2016; Boonstra & de Boer, 2014; Cote & Nightingale, 2012).

Many current approaches to ecosystem-related policies and/or reform have been blind to the historical path-dependency of ecosystem services and their associated values. This has misrepresented

the benefits and burdens associated with the preferences of different groups of people, the “complex mosaics of cultural groups and social classes” (Li, 1996, p. 508) with various and opposed interests, as well as limits to their agency in pursuing alternative ecosystem management or use options (McCusker & Carr, 2006; White & White, 2012).

As well as influencing individual actions, individual values can also be shared through society into more generally held norms and eventually incorporated into institutions that affect collective behaviour (Everard et al., 2016). Such behaviours may include restricted use, protection of certain species or spaces, and modifications of ecosystems that strengthen the generation of a range of benefits, such as agro-ecosystems and their associated biocultural diversity (Berkes & Folke, 1998; Cocks & Wiersum, 2014; Comberti et al., 2015).

Our model emphasises the impact of aspects of well-being supported directly by ecosystem-related benefit bundles. However, broad material, social and relationalⁱⁱ well-being (beyond those dimensions of well-being directly supported by ecosystems)

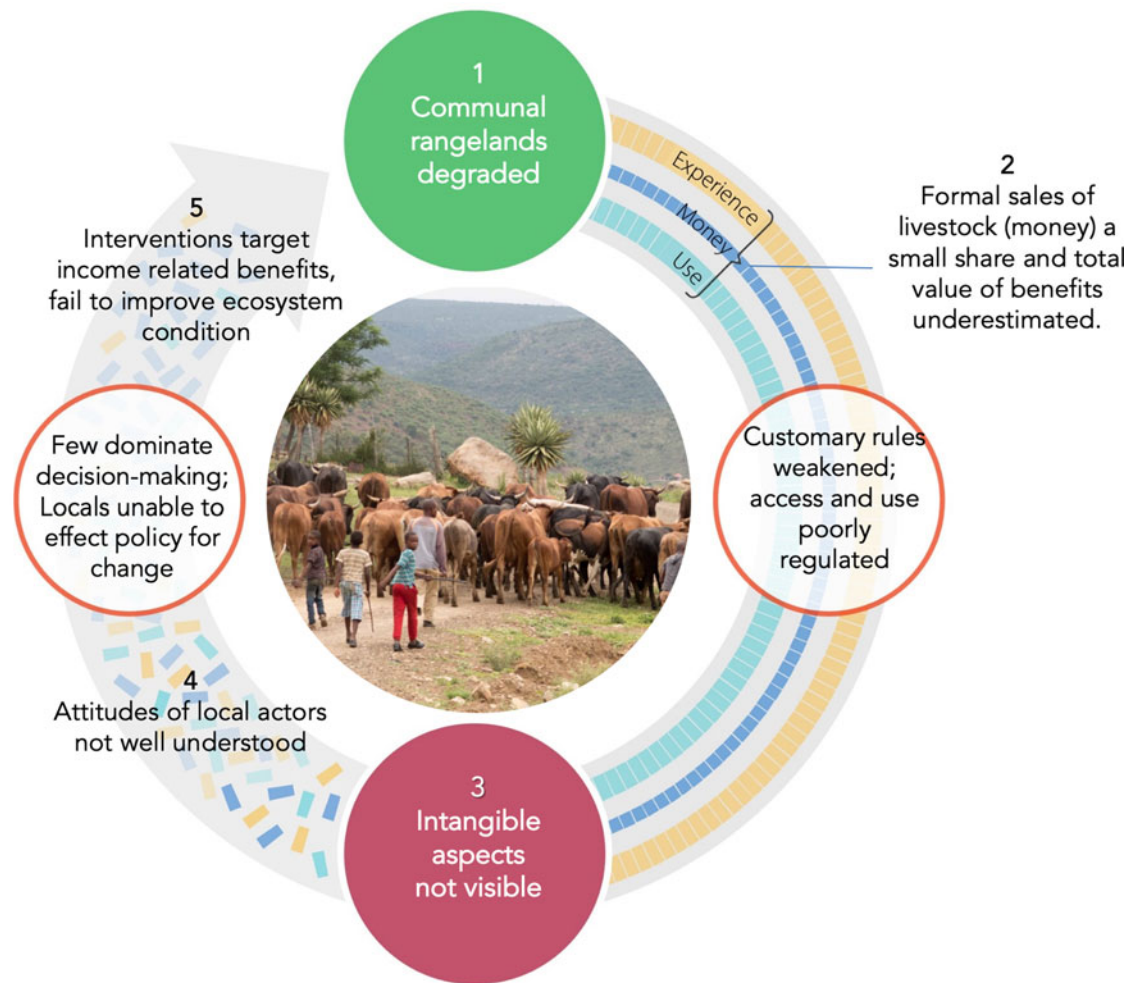


Fig. 2. The conceptual framework as applied to the case study of degraded communal rangelands in southern Africa. The graphics are edited to illustrate how money is a minor part of the benefit basket (2, narrow blue line) and that experience and use are invisible and undervalued (2, yellow and turquoise lines). On the left hand side, the different components of the benefit basket (different coloured pieces) may influence the values and attitudes of local actors (4), but these are not well understood in decision making, and local actors are only involved to a limited degree. Actions to improve rangeland conditions (5) remain focused on reducing stocking rates through sales and thus achieve increased income benefits, but uptake of opportunity for increased sales is low and interventions to change rangeland management and use, often fail. Illustration by J. Lokrantz/Azote. Photograph by Rauri Alcock.

such as income or social status will influence and constrain people's agency to pursue certain courses of action (Sen, 1999), with implications for the environment. For example, a fisher might have positive attitudes towards a marine protected area, but still illegally fish there, in order to sell the catch and meet an acute need to pay their children's school fees. In this example, pro-environmental values are traded-off against other values that are shaped by broader well-being needs.

In summary we can see that a complex interplay of attitudes and values, broader well-being considerations, as well as agency and institutions over time, at individual and collective levels, result in behaviour and actions that may positively or negatively impact ecosystems and the services they provide. These behaviours and how they affect ecosystem dynamics complete the well-being–ecosystem loop. The outcomes of this feedback loop may then be interpreted as reinforcing responsible and sustainable outcomes (which can be termed stewardship) or as degrading ecosystems (and consequently reducing well-being and potentially leading to poverty traps). The conceptual model with its different components, indicated as 1 to 5 in Figure 1, can guide an analysis

of where interventions may shift the feedback towards sustainable outcomes.

Two case studies, one thematic and one regional, illustrate aspects of the conceptual model and its implications for stewardship and overexploitation. While Figure 1 represents a simplified scheme, in reality, different components of our model have their own dynamics, interactions and feedbacks, and our understanding of respective parts is not evenly distributed. This is further illustrated and to some extent explored in the case studies. The case of communal rangeland management (Box 2, Figure 2) suggests that an incomplete understanding of the full benefit basket, how and why it benefits different groups of people unequally, and the constraints on individual agency have hindered successive waves of interventions to improve rangeland condition and outputs. A regional case study from Kalimantan, Indonesia (Box 3, Figure 3), on the conversion of mixed agricultural and natural landscapes to large-scale palm-oil production, illustrates the shifts in benefit baskets and their beneficiaries, and how this has influenced local people's attitudes and agency in land management decisions.

Box 3. Evolving social-ecological relationships in Kalimantan.

In the West Kalimantan province of Indonesia, the landscape is changing fast, with state-driven expansion of oil palm as a rural development strategy resulting in mosaics of diversified agriculture and forest transformed into a sea of monoculture plantations. Oil palm has generated handsome revenues in the form of profits, taxes and fees, but the plantation model is skewed towards benefiting large corporations and elites at the expense of smallholders (Cramb & McCarthy, 2016; Li, 2017). Traditionally, local small-scale Dayak farmers had a diverse relationship with the forest and agricultural land, and tended to derive a **benefits basket** (see point 2 in Figure 3) from multifunctional landscapes, including food security (representing the 'use' mechanism), income security across generations and the flexibility to respond to crises and opportunity (monetary mechanism), as well as the ability to retain autonomy and identity as farmers (experiential mechanism) (De Vos, 2016). This range of benefits was also associated with customary institutions regulating use and access to resources.

However, in the dynamics of this environmental change, the state has co-opted **customary institutions** and changed **rules and norms associated with access**, and in the process widened inequity and led to the loss of access to diverse and flexible farming futures (Li, 2017) (see the red circles in Figure 3). One specific impact of inequity caused by the plantation model is apparent in how the position and livelihood of Dayak women in this already patriarchal community is undermined. Customary tenure systems are a complex, nested mix of communal ownership, as well as ownership based on descent and heritable ownership, where there is no gender differentiation (White & White, 2012). However, the shifting of land tenure from the community to the state and the concomitant practice of registering smallholder plots to the 'family head' has eroded women's customary rights to land (point 3 in Figure 3). At the same time, loss of the recognition of multiple functions of land has led to both men and women smallholders becoming a class of plantation labourer (Li, 2015; White & White, 2012), with very limited agency to influence any land use decisions.

The threat of loss of well-being in the denigration of local rights and multiple functions of the ecosystem has mobilized local and Dayak farmers to actively **express their values** (see 4 in Figure 3) by rejecting and resisting plantation projects throughout West Kalimantan (De Vos, 2016; Morgan, 2017). Rural women who are traditionally excluded from politics have also participated actively in protests against oil palm in an unexpected expression of **agency**. While there have been some successes in the withdrawal of concessions by district heads (De Vos, 2016; Morgan, 2017) the continued expansion of palm throughout the region appears unstoppable. Resistance actions are not universal across all farmers (see 5 in Figure 3). Farmers' assessments of threats (or opportunities) presented by oil palm are rooted in how they perceive variable property and access arrangements (De Vos, 2016; Maharani et al., 2019) and in their aspirations for the future (Li, 2017). The diversity of responses highlights the **complexity of social differentiation and different perceptions of benefits and well-being** derived from the environment (4 in Figure 3). The rural struggles around land and dispossession in West Kalimantan are not driven merely by a switch to a new crop, but relate to the loss of identity and well-being gained from the management of a diverse landscape, and the detachment of local communities from decision-making and crafting rules for managing the landscape.

The State's various attempts to improve cooperation from local farmers have focused mainly on improving laws and policies on plantations, but such actions have failed to recognize the fundamental importance of addressing the loss of multiple ecosystem benefits (including material and non-material) and the effects of shifting power over land to outside companies and elites. A rural development strategy that aims to improve social, economic and environmental sustainability would strive to better engage with the perceptions and values of the local and Dayak farmers, and create opportunities for palm oil cultivations and other crops at a scale that can enable stewardship over ecosystem function. Failure to apply such a strategy can lead to a downward spiral of erosion of multidimensional well-being, stewardship values and capacity, as well as ecosystem functions and diverse benefits.

3. Discussion

This article draws together multiple strands of literature to call for a more holistic and dynamic understanding of the reciprocal relationship between ecosystems and human well-being. This expands on dominant conceptualizations of both ecosystem services, which tend towards a linear and unidirectional relationship of benefits from ecosystems to human well-being, and conservation, which focuses on reducing the negative human impacts on the environment. In contrast, we aim to emphasize the feedbacks occurring as a result of the multiple and reciprocal links between ecosystems and well-being, and specifically how behaviours towards ecosystems may be affected by the way in which ecosystems benefit people, and how these benefits are distributed and perceived.

Through the conceptual model we bring together sets of literature that have previously interacted only to a limited degree. One reason for this may be the different epistemological assumptions and approaches related to the different scales and resolutions on which the respective research traditions focus. We acknowledge epistemological tensions amongst some of the relevant academic traditions brought together in our conceptual framework. For example, psychological research into connections between values,

attitudes and pro-environmental behaviour or action typically focus on the individual, while aggregate valuations of ecosystem services and studies of institutions typically explore patterns and relationships amongst groups or populations, that is, at the collective level. Another example is that while the institutional economics theory of collective action for sustainable governance or management of an ecosystem is rooted in rational choice (Ostrom, 1990), social anthropology and critical institutionalism suggest that motivations are influenced by social concerns, psychological preferences and culturally and historically shaped ideas that are not always rational (Cleaver, 2012). We offer the conceptual model as a starting point for generating further dialogue and new hypotheses. It may also serve as a bridging concept between different 'silos' of understanding (i.e. a bridging concept in the broadest sense of the word – as a tool to foster interdisciplinarity (Baggio et al., 2015; Deppisch & Hasibovic, 2013)).

Our objective for developing the conceptual framework is to enable research and understanding of a more holistic view of interactions between ecosystems and human well-being, using a SES approach. We hope that this can contribute to creating space for multiple framings of the ecosystem and the human



Fig. 3. The conceptual framework as applied to the Kalimantan case study, illustrating the shift from mixed forest and smallholder agriculture towards oil-palm cultivation. The graphics are edited to capture how the shifts lead to loss of benefits related to use for food, materials, etc. and experience of identity and autonomy as farmers (2, narrowing lines of yellow and turquoise), and a change in income security over generations towards dominance of sales of oil palm for money (2, expanding width of blue line). On the left hand side, the blue pieces representing money dominate, but the recognition of the use and experience benefits (yellow and turquoise pieces) may be reflected by local values expressed in community protests. Illustration by J. Lokrantz/Azote. Photograph by Icaro Cooke Viera.

well-being relationship including, but not limited to, an ecosystem service conceptualization. We see our approach as complementary to the other ways of conceptualizing human–nature interactions holistically and dynamically, including in the humanities and indigenous and local knowledge systems, which have often challenged dominant scientific discourses (Caillon *et al.*, 2017; Gavin *et al.*, 2015; Pascual *et al.*, 2017). While not emphasizing well-being, Comberti *et al.* (2015) discuss closing the loop between humans and ecosystems and recognizing how people are involved in actively managing the ecosystem services that are derived from ecosystems, thus recognizing the people's services to nature as well as nature's services to people. They find rich evidence of such a closed feedback loop in many indigenous and local societies, including a range of practices that support ecosystem service generation such as protecting 'useful' species (earthworms, nitrogen-fixing plants, pest-eating birds), that are embedded in rituals and taboos (see also Everard *et al.*, 2016; Whitmarsh, 2009). We view these findings as complementary to our framework and further support for the need to identify barriers and bridges to closing the loop in different contexts.

3.1. Implications for development interventions

There have been multiple calls for a more thorough inclusion of nature and ecosystem dynamics in our understanding of poverty alleviation (Schleicher *et al.*, 2018). Our conceptual model highlights how the mechanisms by which people derive benefits from ecosystems can potentially influence values, attitudes and actions towards these ecosystems that support human well-being. Informed by a SES perspective, this conceptual model compels us to examine the potential impact of poverty alleviation and development interventions on benefit mechanisms and on individual and collective action in relation to ecosystems. Furthermore, it is critical that interventions are aware of and support rather than erode existing values, rules and norms that are congruent with sustainable ecosystem management. For example, customary rules and tenure systems in many African communities revolve around sacred sites, where use and entry are restricted, which protects biodiversity as well as ecosystem functions (Bodin *et al.*, 2006; Sheridan & Nyamweru, 2008; Tengö *et al.*, 2007). Furthermore, the knowledge and management practices of many indigenous communities are embedded in a notion of

reciprocity between people and nature that includes responsible behaviour (Comberti et al., 2015). Interventions that build on such local best practices for sustainability, also have the legitimacy of being community-owned (Mistry et al., 2016; Sterling, 2017b).

The conceptual model encompasses a number of potential points of intervention along the feedback cycle. Firstly, many interventions aim to increase the benefits that people receive from ecosystems (see 1–2 in Figure 1). This may be achieved by, for example, improved access to natural spaces, or skills development to increase yields or catch, or better connections to markets at which ecosystem products can be sold. While these strategies have been effective, we suggest that such interventions could be further improved by explicitly considering the multitude of different benefit mechanisms (direct use, monetary and experience), as well as who can access the benefits and who bears the costs. Secondly, interventions may aim to improve awareness of the full range of benefits that are derived from an ecosystem (see 2–4, Figure 1), as this may have a positive influence on values of responsibility and care, which in turn play an important role in promoting stewardship of nature. Thirdly, interventions can support the agency of individuals, collectives and communities to act on their values (4–5, Figure 1) by, for example, recognizing territorial rights, indigenous knowledge and communal/customary governance systems (Caillon et al., 2017; Comberti et al., 2015; Sterling et al., 2017a). While all of these strategies are typical conservation and development strategies, our conceptual model highlights the inherent interdependencies between these. For example, efforts to support community conservation of ecosystems are common, but may be blind to changing or inequitable benefits from ecosystems that could undermine the values and incentives for community members to invest in conservation or stewardship actions. Similarly, development interventions to improve the management and productivity of rangelands (e.g. the introduction of improved breeding stock linked to marketing schemes) typically target and materially benefit the more affluent and educated owners of cattle, who are a minority (Vetter, 2013). The outcomes of these interventions may negatively impact the benefit basket for marginalized community members such as women and the youth, for example, preventing their access to experiential benefits from these rangelands. This, in turn, affects the relationships that these vulnerable groups have with their local ecosystems, reducing both their local ecological knowledge and their incentives for stewardship (Box 2). The conceptual model could be used as a diagnostic tool to identify points in the feedback loop where interventions can unlock the potential for broader support of stewardship.

In particular, we emphasize the mechanism of experience which contributes to aspects of well-being that are commonly overlooked in economically focused development strategies (Büscher & Dressler, 2007; Masterson et al., 2017; Russell et al., 2013). Insights from the humanities and social sciences illustrate that experiences in nature nurture a sense of care and responsibility towards the ecosystem (Bennett et al., 2018; Bramston et al., 2011; Chawla, 2009; Enqvist, 2017; Giusti et al., 2014; Raghuram, 2016). Development interventions that invest in sustaining existing experiential activities and benefits could both improve neglected aspects of well-being (such as relational well-being), and maintain and encourage protective actions and sustainable management practices of the ecosystem, thus sustaining a virtuous cycle (positive feedback loop) that ultimately supports material benefits from the ecosystem in the long term. This is particularly important in the face of development trajectories that commonly favour ecosystem simplification, commercialization and urbanization. This is evident

in the case of Kalimantan (Box 3), where homogenous oil palm plantations are promoted as a form of rural development without understanding the local systems of value around farmer identity, land management and tenure, as well as diverse perceptions of well-being. Explicit consideration of experiential ecosystem benefits, and the potential trade-offs with direct use and money, challenges dominant ideas in development practice and poverty alleviation. Focusing on experiential benefits is key to maintaining the existing ties and positive relations between people and nature, which is important in motivating and maintaining local stewardship actions (Giusti et al., 2014; Krasny & Tidball, 2012). It may also serve as an entry point to better understand the cultural dimensions of experiencing as well as managing ecosystems, which many have argued should inform the implementation and evaluation of poverty alleviation and development initiatives (Díaz et al., 2018; Schleicher et al., 2018; Sterling et al., 2017a).

There is a need for deeper alignment between processes of use, monetary benefits and experiences, as well as institutions and agency to support sustainable development and stewardship. Both case studies (Box 2 & 3) call for a broader and more inclusive range of benefit mechanisms to be recognized, and challenge dominant discourses that inform and justify the types of intervention approaches and institutional structures that empower large corporations and elites over small-scale landholders. Such a shift to a broader perspective would require a transformation of development interventions at a global (conceptual) and at the local (implementation) level; by extending emphasis beyond material needs, and incorporating subjective as well as relational aspects of well-being for communities and individuals (Caillon et al., 2017; Sterling et al., 2017a). Such an approach also emphasizes the need for interventions to be relevant for the local context, and for more representative and inclusive local institutions that can represent the diverse range of values held by communities.

3.2. Key uncertainties and gaps

Figure 1 suggests a distinct feedback loop, but there remain a number of key uncertainties on how or whether such a feedback operates in different contexts, which needs further elaboration and research from different perspectives. The influence of perceived benefits from nature on attitudes and values is complex and uncertain and it is difficult to attribute actions and behavioural outcomes to just one aspect of people's lives. Firstly, not all benefits may be perceived. For example, supporting or regulating ecosystem services are likely to be taken for granted unless they decline, and typically receive less emphasis in participatory assessments of ecosystem services (Martín-López et al., 2012; Rodríguez et al., 2006). Meanwhile, management institutions may fail to recognize benefits even if individuals are aware of them (see also Box 2). For example, capitalist markets generally fail to recognize freely provided ecosystem services that have no price (Daily et al., 2009; Vogl et al., 2017), while governance institutions may overlook, undervalue or choose to ignore direct or experience-oriented benefits to marginalized communities (Reed et al., 2017). Indeed, even when benefits are perceived, this awareness may not be enough to change behaviour towards, for example, stewardship of ecosystem services (Blake, 1999; Kollmuss & Agyeman, 2002). Furthermore, recent work has highlighted the plurality of values and how contextual values can be formed by deliberative valuation exercises (Kenter, 2016). Context-specific values and attitudes that are relevant for stewardship actions therefore likely require benefits to be consciously considered and reflected upon. Finally, it is unclear to what extent

values towards ecosystems are necessarily influenced by the perception of benefits from ecosystems, as, for example, various value systems have notions of care and responsibility that are not dependent on utilitarian reciprocity with ecosystems (Caillon *et al.*, 2017; Cocks *et al.*, 2012; Cooper *et al.*, 2016). Thus, there is a clear need for collaborative research across different disciplines to better tease out the pathways from experience and perceptions of the benefit basket (2–3; Figure 1), to attitudes and values (4; Figure 1) and the implications for actions (5; Figure 1), and finally to impact on ecosystem capacity to generate benefits (1; Figure 1).

Another key gap is in our understanding of how the different ways people benefit from ecosystems (e.g. money, use, experience) can influence attitudes and actions differently, i.e., do the benefits of, e.g., use and experience result in a different set of attitudes and actions than monetary benefits? For example, while women's perceptions of the forest as contributing to their material well-being may be relatively well documented, much less is known about how they feel about, and value, the forest where they harvest firewood. Still less is understood about the contribution of use and experiential benefits to how these feelings and values are developed and mobilized towards action, for example, on behalf of the environment. To what extent are women who derive psychological and relational well-being from collecting firewood (Cocks *et al.*, 2012) inclined to and able to act to protect forest resources? And how does the loss or gain of such aspects of well-being influence behaviour? Along similar lines, some critiques of payments for ecosystem services (PES) schemes have highlighted that monetary benefits from ecosystems may 'crowd out' less instrumental values of stewardship based on experiential connections with nature (Ezzine-De-Blas *et al.*, 2015). Our conceptual model illustrates the importance of gaining a better understanding of these interactions by highlighting the variety of mechanisms through which ecosystems confer benefits, as well as the different ways in which these benefits then shape attitudes and behaviours.

3.3. Future research directions and conclusions

As a potential bridging object between different bodies of knowledge, our conceptual model has helped us to define some critical interdisciplinary research questions for understanding how feedbacks between ecosystems and well-being can nurture or erode stewardship. The first set of questions addresses how different mechanisms of benefit influence values and behaviour. For example, does monetization change attitudes, values and behaviours, and what are the implications for longer term environmental sustainability? How strongly do experiential mechanisms influence action relative to monetary benefits and direct use? Additionally, how do different benefit mechanisms encourage behaviours with effects at different scales (e.g. in one's community vs. at a national level)? A second avenue for future research concerns the interplay between perceived general well-being and the aspects of well-being directly supported by ecosystem services: what is the relative importance of the contribution of ecosystem service-related aspects of well-being to stewardship actions, as opposed to the contribution of broader measures of well-being? We would particularly like to call for research designs that assess components of the loop, while also acknowledging a holistic perspective and the dynamics of the full feedback loop.

A third avenue concerns the role of equity in ecosystem benefits and the ultimate influence of socio-economic inequalities on ecosystem sustainability (Hamann *et al.*, 2018). A growing body of research has argued for the inclusion of social feedbacks,

where perceptions of equity and justice can significantly influence ecosystem governance (or stewardship) (e.g. Bodin *et al.*, 2006; Pascual *et al.*, 2014). Loft *et al.* (2017) and McElwee *et al.* (2014) show how PES policies in Vietnam are reshaped by rural actors to better reflect local notions of equity, born of deep-rooted place-specific social and cultural norms. While there is a large body of literature that assesses the effect of economic inequality on collaboration through, for example, behavioural experiments, fewer studies assess the links between equity and cooperation in terms of managing an environmental resource (Baland *et al.*, 2007; Cardenas, 2003). There is a need for systematic evidence of whether perceived equity (in terms of participation, access to resources and distribution of benefits) leads to stewardship actions and outcomes in the real world. This is an important area for future research and can help to illustrate how these feedbacks have influence on the linkages between well-being and ecosystem stewardship as illustrated in the conceptual model.

Maintaining or enhancing the contributions of ecosystems to human well-being while ensuring that these ecosystems continue to provide benefits for future generations is a major global challenge. How our proposed feedback loop operates within the increasing scope and scale of environmental and social change, including climate change, large scale human migrations and ecosystem degradation could be critical in addressing future sustainability challenges. Rather than considering relationships between ecosystems and well-being as unidirectional, and linear, this paper presents a conceptual model to capture a more dynamic interplay and feedback between well-being and stewardship of nature. Furthering our understanding of the different ways through which well-being can be derived from interactions with ecosystems, as well as the processes through which well-being in turn shapes values and behaviour towards the environment, can highlight potential points of intervention that may lead to virtuous cycles with positive outcomes for both human well-being and ecosystem health through the promotion of environmental stewardship.

Notes

ⁱ <http://www.pecs-science.org/>

ⁱⁱ The relational dimension of well-being refers to social connections, power and identity, and the relationships that a person must be able to enter into in order to continue meeting the needs that are important for their well-being (McGregor, 2007).

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References

- Ainslie, A. (2005). Farming cattle, cultivating relationships: cattle ownership and cultural politics in Peddie District, Eastern Cape. *Social Dynamics*, 31(1), 129–156. <https://doi.org/10.1080/02533950508628699>
- Ajzen, I., & Fishbein, M. (1980). *Understanding Attitudes and Predicting Social Behaviour*. Englewood Cliffs, USA: Prentice-Hall.
- Baggio, J. A., Brown, K., & Hellebrandt, D. (2015). Boundary object or bridging concept? A citation network analysis of resilience. *Ecology and Society*, 20(2), 2. <https://doi.org/10.5751/ES-07484-200202>
- Baland, J.-M., Bardhan, P. K., & Bowles, S. (2007). *Inequality, Cooperation, and Environmental Sustainability*. New York, USA: Russell Sage Foundation.
- Barbier, E. B. (2010). Poverty, development, and environment. *Environment and Development Economics*, 15(6), 635–660. <https://doi.org/10.1017/S1355770X1000032X>
- Bennett, N. J., Whitty, T. S., Finkbeiner, E., Pittman, J., Bassett, H., Gelcich, S., & Allison, E. H. (2018). Environmental stewardship: a conceptual review and analytical framework. *Environmental Management*, 61(4), 597–614. <https://doi.org/10.1007/s00267-017-0993-2>
- Berbés-Blázquez, M., Bunch, M. J., Mulvihill, P. R., Peterson, G. D., & van Wendel de Joode, B. (2017). Understanding how access shapes the transformation of ecosystem services to human well-being with an example from Costa Rica. *Ecosystem Services*, 28, 320–327. <https://doi.org/10.1016/j.ecoser.2017.09.010>
- Berbés-Blázquez, M., González, J. A., & Pascual, U. (2016). Towards an ecosystem services approach that addresses social power relations. *Current Opinion in Environmental Sustainability*, 19, 134–143. <https://doi.org/10.1016/j.cosust.2016.02.003>
- Berkes, F., Colding, J., & Folke, C. (eds) (2003). *Navigating Social-Ecological Systems: Building Resilience for Complexity and Change*. Cambridge, UK: Cambridge University Press.
- Berkes, F., & Folke, C. (eds) (1998). *Linking Social and Ecological Systems: Management Practices and Social Mechanisms for Building Resilience*. Cambridge, UK: Cambridge University Press.
- Blake, J. (1999). Overcoming the 'value-action gap' in environmental policy: Tensions between national policy and local experience. *Local Environment*, 4(3), 257–278. <https://doi.org/10.1080/13549839908725599>
- Bodin, O., Tengö, M., Norman, A., Lundberg, J., & Elmqvist, T. (2006). The value of small size: loss of forest patches and ecological thresholds in southern Madagascar. *Ecological Applications*, 16(2), 440–451. [https://doi.org/10.1890/1051-0761\(2006\)016\[0440:TVOSL\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2006)016[0440:TVOSL]2.0.CO;2)
- Boonstra, W. J., & de Boer, F. W. (2014). The historical dynamics of social-ecological traps. *Ambio*, 43(3), 260–274. <https://doi.org/10.1007/s13280-013-0419-1>
- Bramston, P., Pretty, G., & Zammit, C. (2011). Assessing environmental stewardship motivation. *Environment and Behavior*, 43(6), 776–788. <https://doi.org/10.1177/0013916510382875>
- Breslow, S. J., Sojka, B., Barnea, R., Basurto, X., Carothers, C., Charnley, S., ... Levin, P. S. (2016). Conceptualizing and operationalizing human wellbeing for ecosystem assessment and management. *Environmental Science & Policy*. <https://doi.org/10.1016/j.envsci.2016.06.023>
- Brown, K., & Westaway, E. (2011). Agency, capacity, and resilience to environmental change: lessons from human development, well-being, and disasters. *Annual Review of Environment and Resources*, 36(1), 321–342. <https://doi.org/10.1146/annurev-environ-052610-092905>
- Büscher, B., & Dressler, W. (2007). Linking neoprotectionism and environmental governance: On the rapidly increasing tensions between actors in the environment-development nexus. *Conservation and Society*, 5(4), 586–611.
- Caillon, S., Cullman, G., Verschuuren, B., & Sterling, E. J. (2017). Moving beyond the human-nature dichotomy through biocultural approaches: including ecological well-being in resilience indicators. *Ecology and Society*, 22(4), 27. <https://doi.org/10.5751/ES-09746-220427>
- Cardenas, J. C. (2003). Real wealth and experimental cooperation: Experiments in the field lab. *Journal of Development Economics*, 70(2), 263–289. [https://doi.org/10.1016/S0304-3878\(02\)00098-6](https://doi.org/10.1016/S0304-3878(02)00098-6)
- Casaló, L. V., & Escario, J.-J. (2018). Heterogeneity in the association between environmental attitudes and pro-environmental behavior: A multilevel regression approach. *Journal of Cleaner Production*, 175, 155–163. <https://doi.org/10.1016/J.JCLEPRO.2017.11.237>
- Chaigneau, T. (2013). *Understanding Community Support Towards Three Marine Protected Areas in the Visayas Region of the Philippines* (University of East Anglia). Retrieved from https://ueaeprints.uea.ac.uk/48083/1/Tomas_Chaigneau-PhD_Thesis_Post_Viva_131205.pdf
- Chan, K. M. A., Balvanera, P., Benessaiah, K., Chapman, M., Diaz, S., Gómez-Baggeth, E., ... Turner, N. (2016). Opinion: Why protect nature? Rethinking values and the environment. *Proceedings of the National Academy of Sciences*, 113(6), 1462–1465. <https://doi.org/10.1073/pnas.1525002113>
- Chan, K. M. A., Satterfield, T., & Goldstein, J. (2012). Rethinking ecosystem services to better address and navigate cultural values. *Ecological Economics*, 74, 8–18. <https://doi.org/10.1016/j.ecolecon.2011.11.011>
- Chapin, F. S., Folke, C., & Kofinas, G. P. (2009). A framework for understanding change. In F. S. Chapin III, G. P. Kofinas, & C. Folke (eds), *Principles of Ecosystem Stewardship: Resilience-Based Natural Resource Management in a Changing World* (pp. 3–28). https://doi.org/10.1007/978-0-387-73033-2_1
- Chawla, L. (2009). Growing up green: becoming an agent of care for the natural world. *Journal of Development Processes*, 4(1), 6–23.
- Cleaver, F. (2012). *Development Through Bricolage: Rethinking Institutions for Natural Resource Management*. Abingdon, Oxon: Routledge.
- Cocks, M., Alexander, J., Mogano, L., & Vetter, S. (2016). Ways of belonging: meanings of "nature" among Xhosa-speaking township residents in South Africa. *Journal of Ethnobiology*, 36(4), 820–841. <https://doi.org/10.2993/0278-0771-36.4.820>
- Cocks, M., Dold, T., & Vetter, S. (2012). 'God is my forest' – Xhosa cultural values provide untapped opportunities for conservation. *South African Journal of Science*, 108(5/6), 1–8.
- Cocks, M., & Wiersum, F. (2014). Reappraising the concept of biocultural diversity: a perspective from South Africa. *Human Ecology*, 42(5), 727–737. <https://doi.org/10.1007/s10745-014-9681-5>
- Comberti, C., Thornton, T. F., Wyllie de Echeverria, V., & Patterson, T. (2015). Ecosystem services or services to ecosystems? Valuing cultivation and reciprocal relationships between humans and ecosystems. *Global Environmental Change*, 34, 247–262. <https://doi.org/10.1016/j.gloenvcha.2015.07.007>
- Cooper, N., Brady, E., Steen, H., & Bryce, R. (2016). Aesthetic and spiritual values of ecosystems: Recognising the ontological and axiological plurality of cultural ecosystem 'services.' *Ecosystem Services*, 21, 218–229. <https://doi.org/10.1016/J.ECOSER.2016.07.014>
- Cote, M., & Nightingale, A. J. (2012). Resilience thinking meets social theory: Situating social change in socio-ecological systems (SES) research. *Progress in Human Geography*, 36(4), 475–489. <https://doi.org/10.1177/0309132511425708>
- Coulthard, S. (2012). Can we be both resilient and well, and what choices do people have? Incorporating agency into the resilience debate from a fisheries perspective. *Ecology and Society*, 17(1), 4. <https://doi.org/10.5751/ES-04483-170104>
- Cramb, R. A., & McCarthy, J. F. (eds) (2016). *The Oil Palm Complex: Smallholders, Agribusiness and the State in Indonesia and Malaysia*. Singapore: National University of Singapore Press.
- Cruz-García, G. S., Sachet, E., Blundo-Canto, G., Vanegas, M., & Quintero, M. (2017). To what extent have the links between ecosystem services and human well-being been researched in Africa, Asia, and Latin America? *Ecosystem Services*, 25, 201–212. <https://doi.org/10.1016/j.ecoser.2017.04.005>
- Daily, G. C., Polasky, S., Goldstein, J., Kareiva, P. M., Mooney, H. A., Pejchar, L., ... Shallenberger, R. (2009). Ecosystem services in decision

- making: time to deliver. *Frontiers in Ecology and the Environment*, 7(1), 21–28. <https://doi.org/10.1890/080025>
- Daw, T., Brown, K., Rosendo, S., & Pomeroy, R. (2011). Applying the ecosystem services concept to poverty alleviation: the need to disaggregate human well-being. *Environmental Conservation*, 38(4), 370–379. <https://doi.org/10.1017/S0376892911000506>
- De Vos, R. E. (2016). Multi-functional lands facing oil palm monocultures: a case study of a land conflict in West Kalimantan, Indonesia. *ASEAS – Austrian Journal of South-East Asian Studies*, 9(1), 11–32. <https://doi.org/10.14764/10.ASEAS-2016.1-2>
- Deppisch, S., & Hasibovic, S. (2013). Social-ecological resilience thinking as a bridging concept in transdisciplinary research on climate-change adaptation. *Natural Hazards*, 67(1), 117–127. <https://doi.org/10.1007/s11069-011-9821-9>
- Díaz, S., Demissew, S., Carabias, J., Joly, C., Lonsdale, M., Ash, N., ... Zlatanova, D. (2015). The IPBES Conceptual Framework – connecting nature and people. *Current Opinion in Environmental Sustainability*, 14, 1–16. <https://doi.org/10.1016/j.cosust.2014.11.002>
- Díaz, S., Pascual, U., Stenseke, M., Martín-López, B., Watson, R. T., Molnár, Z., ... Shirayama, Y. (2018). Assessing nature's contributions to people: Recognizing culture, and diverse sources of knowledge, can improve assessments. *Science*, 359(6373), 270–272. <https://doi.org/10.1126/science.aap8826>
- Diener, E. (1984). Subjective well-being. *Psychological Bulletin*, 95(3), 542–575.
- Diener, E., & Suh, E. M. (eds) (2000). *Culture and Subjective Well-Being*. Cambridge, MA: MIT Press.
- Dietz, T., Fitzgerald, A., & Shwom, R. (2005). Environmental values. *Annual Review of Environment and Resources*, 30(1), 335–372. <https://doi.org/10.1146/annurev.energy.30.050504.144444>
- Djoudi, H., & Brockhaus, M. (2011). Is adaptation to climate change gender neutral? Lessons from communities dependent on livestock and forests in northern Mali. *International Forestry Review*, 13(2), 123–135. <https://doi.org/10.1505/146554811797406606>
- Doyal, L., & Gough, I. (1991). *A Theory of Human Need*. London, UK: Palgrave. <https://doi.org/10.1007/978-1-349-21500-3>
- Eagly, A. H., & Chaiken, S. (1993). *The Psychology of Attitudes*. Fort Worth, TX: Harcourt Brace Jovanovich College Publishers.
- Enqvist, J. (2017). *Stewardship in an Urban World: Civic Engagement and Human-Nature Relations in the Anthropocene*. Stockholm: Stockholm Resilience Centre, Stockholm University.
- Escobar, A. (1996). Constructing nature: elements for a poststructural political ecology. In R. Peet & M. Watts (eds), *Liberation Ecologies: Environment, Development, Social Movements* (pp. 46–68). London, UK: Routledge.
- Everard, M., Reed, M. S., & Kenter, J. O. (2016). The ripple effect: Institutionalising pro-environmental values to shift societal norms and behaviours. *Ecosystem Services*, 21, 230–240. <https://doi.org/10.1016/j.ecoser.2016.08.001>
- Ezzine-De-Blas, D., Corbera, E., & Lapeyre, R. (2015). Crowding-in or crowding-out? A conceptual framework to understand motivations in payments for ecosystem services. *Resource Politics 2015*. Brighton: Institute of Development Studies.
- Folke, C., Biggs, R., Norström, A. V., Reyers, B., & Rockström, J. (2016). Social-ecological resilience and biosphere-based sustainability science. *Ecology and Society*, 21(3), 41. <https://doi.org/10.5751/ES-08748-210341>
- Gavin, M. C., McCarter, J., Mead, A., Berkes, F., Stepp, J. R., Peterson, D., & Tang, R. (2015). Defining biocultural approaches to conservation. *Trends in Ecology and Evolution*, 30(3), 140–145. <https://doi.org/10.1016/j.tree.2014.12.005>
- Giusti, M., Barthel, S., & Marcus, L. (2014). Nature routines and affinity with the biosphere: a case study of preschool children in Stockholm. *Children, Youth and Environments*, 24(3), 16. <https://doi.org/10.7721/chilyoutenvi.24.3.0016>
- Griggs, D., Stafford-Smith, M., Gaffney, O., Rockström, J., Ohman, M. C., Shyamsundar, P., ... Noble, I. (2013). Policy: sustainable development goals for people and planet. *Nature*, 495(7441), 305–307. <https://doi.org/10.1038/495305a>
- Guerry, A. D., Polasky, S., Lubchenco, J., Chaplin-Kramer, R., Daily, G. C., Griffin, R., ... Vira, B. (2015). Natural capital and ecosystem services informing decisions: From promise to practice. *Proceedings of the National Academy of Sciences*, 112(24), 7348–7355. <https://doi.org/10.1073/pnas.1503751112>
- Haider, L. J., Boonstra, W. J., Peterson, G. D., & Schlüter, M. (2018). Traps and sustainable development in rural areas: a review. *World Development*, 101, 311–321. <https://doi.org/10.1016/J.WORLDDEV.2017.05.038>
- Hajer, M. A. (1995). *The Politics of Environmental Discourse: Ecological Modernization and the Policy Process*. Oxford: Oxford University Press.
- Hamann, M., Berry, K., Chaigneau, T., Curry, T., Heilmayr, R., Henriksson, P. J. G., ... Wu, T. (2018). Inequality and the biosphere. *Annual Review of Environment and Resources*, 43(1), 61–83. <https://doi.org/10.1146/annurev-environ-102017-025949>
- Hebinck, P., & Shackleton, C. (2010). *Reforming Land and Resource Use in South Africa: Impact on Livelihoods*. Retrieved from https://www.google.com/books?hl=en&lr=&id=NiwwCgAAQBAJ&oi=fnd&pg=PP1&ots=vtrtN_vhy5&sig=QP7NEERROBT6CMeYvd3vBMXLMcI
- Hogg, M. A., & Vaughan, G. M. (2005). *Social Psychology (4th edition)*. Retrieved from <http://psycnet.apa.org/record/1995-99018-000>
- Homer, P. M., & Kahle, L. R. (1988). A structural equation test of the value-attitude-behavior hierarchy. *Journal of Personality and Social Psychology*, 54(4), 638–646. <https://doi.org/10.1037/0022-3514.54.4.638>
- Jones, N. A., Shaw, S., Ross, H., Witt, K., & Pinner, B. (2016). The study of human values in understanding and managing social-ecological systems. *Ecology and Society*, 21(1), 15. <https://doi.org/10.5751/ES-07977-210115>
- Kenter, J. O. (2016). Editorial: Shared, plural and cultural values. *Ecosystem Services*, 21, 175–183. <https://doi.org/10.1016/j.ecoser.2016.10.010>
- Kepe, T., McGregor, G., & Irvine, P. (2015). Rights of “passage” and contested land use: Gendered conflict over urban space during ritual performance in South Africa. *Applied Geography*, 57, 91–99. <https://doi.org/10.1016/j.apgeog.2014.12.021>
- Klöckner, C. A. (2013). A comprehensive model of the psychology of environmental behaviour – A meta-analysis. *Global Environmental Change*, 23(5), 1028–1038. <https://doi.org/10.1016/j.gloenvcha.2013.05.014>
- Kollmuss, A., & Agyeman, J. (2002). Mind the gap: Why do people act environmentally and what are the barriers to pro-environmental behavior? *Environmental Education Research*, 8(3), 239–260. <https://doi.org/10.1080/13504620220145401>
- Krasny, M. E., & Tidball, K. G. (2012). Civic ecology: a pathway for Earth Stewardship in cities. *Frontiers in Ecology and the Environment*, 10(5), 267–273. <https://doi.org/10.1890/110230>
- Leach, M., Mearns, R., & Scoones, I. (1999). Environmental entitlements: dynamics and institutions in community based natural resource management. *World Development*, 21(2), 225–247. [https://doi.org/10.1016/S0305-750X\(98\)00141-7](https://doi.org/10.1016/S0305-750X(98)00141-7)
- Li, T. (1996). Images of community: discourse and strategy in property relations. *Development and Change*, 27(3), 501–527.
- Li, T. (2015). *Social Impacts of Oil Palm in Indonesia: A Gendered Perspective from West Kalimantan*. Occasional Paper 124. Bogor, Indonesia: CIFOR.
- Li, T. (2017). Intergenerational displacement in Indonesia's oil palm plantation zone. *Journal of Peasant Studies*, 44(6), 1160–1178. <https://doi.org/10.1080/03066150.2017.1308353>
- Loft, L., Le, D. N., Pham, T. T., Yang, A. L., Tjajadi, J. S., & Wong, G. Y. (2017). Whose equity matters? National to local equity perceptions in Vietnam's payments for forest ecosystem services scheme. *Ecological Economics*, 135, 164–175. <https://doi.org/10.1016/J.ECOLECON.2017.01.016>
- Loomis, D. K., & Paterson, S. K. (2014). The human dimensions of coastal ecosystem services: Managing for social values. *Ecological Indicators*, 44, 6–10. <https://doi.org/10.1016/J.ECOLIND.2013.09.035>
- MA (2005). *Ecosystems and human well-being: synthesis*. In *Millennium Ecosystem Assessment* (p. 136). Washington D.C., USA: Island Press. Retrieved from <http://www.millenniumassessment.org/documents/document.356.aspx.pdf>
- Maharani, C. D., Moeliono, M., Wong, G. Y., Brockhaus, M., Carmenta, R., & Kallio, M. (2019). Development and equity: A gendered inquiry in a swidden landscape. *Forest Policy and Economics*, 101, 120–128. <https://doi.org/10.1016/j.forpol.2018.11.002>
- Martín-López, B., Iniesta-Arandia, I., García-Llorente, M., Palomo, I., Casado-Arzuaga, I., Del Amo, D. G., ... Montes, C. (2012). Uncovering

- ecosystem service bundles through social preferences. *PLoS ONE*, 7(6), e38970. <https://doi.org/10.1371/journal.pone.0038970>
- Masterson, V., Tengö, M., & Spierenburg, M. (2017). Competing place meanings in complex landscapes: a social–ecological approach to unpacking community conservation outcomes on the Wild Coast, South Africa. *Society & Natural Resources*, 30(12), 1442–1457. <https://doi.org/10.1080/08941920.2017.1347975>
- McCusker, B., & Carr, E. R. (2006). The co-production of livelihoods and land use change: Case studies from South Africa and Ghana. *Geoforum*, 37, 790–804. <https://doi.org/10.1016/j.geoforum.2005.09.007>
- McElwee, P., Nghiem, T., Le, H., Vu, H., & Tran, N. (2014). Payments for environmental services and contested neoliberalisation in developing countries: A case study from Vietnam. *Journal of Rural Studies*, 36, 423–440. <https://doi.org/10.1016/j.jrurstud.2014.08.003>
- McGregor, J. A. (2007). Researching human wellbeing: from concepts to methodology. In I. Gough & J. A. McGregor (eds), *Well-being in Developing Countries. From Theory to Research* (pp. 316–350). Cambridge, UK: Cambridge University Press.
- Mistry, J., Berardi, A., Tschirhart, C., Bignante, E., Haynes, L., Benjamin, R., ... de Ville, G. (2016). Community owned solutions: identifying local best practices for social-ecological sustainability. *Ecology and Society*, 21(2), 42. <https://doi.org/10.5751/ES-08496-210242>
- Morgan, M. (2017). Women, gender and protest: contesting oil palm plantation expansion in Indonesia. *The Journal of Peasant Studies*, 44(6), 1177–1196. <https://doi.org/10.1080/03066150.2017.1300579>
- North, D. (1991). Institutions. *Journal of Economic Perspectives*, 5(1), 97–112. <https://doi.org/10.1257/jep.5.1.97>
- Ostrom, E. (1990). *Governing the Commons: the Evolution of Institutions for Collective Action*. New York, USA: Cambridge University Press.
- Palomo, I., Felipe-Lucia, M. R., Bennett, E. M., Martín-López, B., & Pascual, U. (2016). Disentangling the pathways and effects of ecosystem service co-production. *Advances in Ecological Research*, 54, 245–283. <https://doi.org/10.1016/bs.aecr.2015.09.003>
- Pascua, P., McMillen, H., Ticktin, T., Vaughan, M., & Winter, K. B. (2017). Beyond services: A process and framework to incorporate cultural, genealogical, place-based, and indigenous relationships in ecosystem service assessments. *Ecosystem Services*, 26, 465–475. <https://doi.org/10.1016/j.ecoser.2017.03.012>
- Pascual, U., Balvanera, P., Diaz, S., Pataki, G., Roth, E., Stenseke, M., ... Wickson, F. (2017). Valuing nature's contributions to people: The IPBES approach. *Current Opinion in Environmental Sustainability*, 26, 7–16. <https://doi.org/10.1016/j.cosust.2016.12.006>
- Pascual, U., Phelps, J., Garmendia, E., Brown, K., Corbera, E., Martin, A., ... Muradian, R. (2014). Social equity matters in payments for ecosystem services. *BioScience*, 64(11), 1027–1036. <https://doi.org/10.1093/biosci/biu146>
- Raghuram, P. (2016). Locating care ethics beyond the Global North. *ACME: An International Journal for Critical Geographies*, 15(3), 511–533. Retrieved from <https://acme-journal.org/index.php/acme/article/view/1353>
- Reed, M. S., Allen, K., Attlee, A., Dougill, A. J., Evans, K. L., Kenter, J. O., ... Whittingham, M. J. (2017). A place-based approach to payments for ecosystem services. *Global Environmental Change*, 43, 92–106. <https://doi.org/10.1016/j.gloenvcha.2016.12.009>
- Reyers, B., Biggs, R., Cumming, G. S., Elmquist, T., Hejnowicz, A. P., & Polasky, S. (2013). Getting the measure of ecosystem services: a social–ecological approach. *Frontiers in Ecology and the Environment*, 11(5), 268–273. <https://doi.org/10.1890/120144>
- Ribot, J. J. C., & Peluso, N. L. N. (2003). A theory of access. *Rural Sociology*, 68(2), 153–181. <https://doi.org/10.1111/j.1549-0831.2003.tb00133.x>
- Rodríguez, J. P., Beard Jr., T. D., Bennett, E. M., Cumming, G. S., Cork, S. J., Agard, J., ... Peterson, G. D. (2006). Trade-offs across space, time, and ecosystem services. *Ecology and Society*, 11(1), 28. <https://doi.org/10.2307/26267786>
- Rokeach, M. (1973). *The Nature of Human Values*. New York, NY, USA: The Free Press.
- Rounsevell, M. D. A., Dawson, T. P., & Harrison, P. A. (2010). A conceptual framework to assess the effects of environmental change on ecosystem services. *Biodiversity and Conservation*, 19(10), 2823–2842. <https://doi.org/10.1007/s10531-010-9838-5>
- Russell, R., Guerry, A. D., Balvanera, P., Gould, R. K., Basurto, X., Chan, K. M. A., ... Tam, J. (2013). Humans and nature: how knowing and experiencing nature affect well-being. *Annual Review of Environment and Resources*, 38(1), 473–502. <https://doi.org/10.1146/annurev-environ-012312-110838>
- Sachs, J. D. (2012). From millennium development goals to sustainable development goals. *Lancet*, 379(9832), 2206–2211. [https://doi.org/10.1016/S0140-6736\(12\)60685-0](https://doi.org/10.1016/S0140-6736(12)60685-0)
- Satterfield, T. (2001). In search of value literacy: suggestions for the elicitation of environmental values. *Environmental Values*, 10(3), 331–359. <https://doi.org/10.3197/096327101129340868>
- Schleicher, J., Schaafsma, M., Burgess, N. D., Sandbrook, C., Danks, F., Cowie, C., & Vira, B. (2018). Poorer without it? The neglected role of the natural environment in poverty and wellbeing. *Sustainable Development*, 26(1), 83–98. <https://doi.org/10.1002/sd.1692>
- Schwartz, S. H. (1994). Are there universal aspects in the structure and contents of human values? *Journal of Social Issues*, 50(4), 19–45. <https://doi.org/10.1111/j.1540-4560.1994.tb01196.x>
- Sen, A. (1999). *Development as Freedom*. New York, USA: Knopf.
- Shackleton, C. M., Shackleton, S. E., & Cousins, B. (2001). The role of land-based strategies in rural livelihoods: the contribution of arable production, animal husbandry and natural resource harvesting in communal areas in South Africa. *Development Southern Africa*, 18(5), 581–604. <https://doi.org/10.1080/0376835012009744>
- Shackleton, S., Campbell, B., Lotz-Sisitka, H., & Shackleton, C. (2008). Links between the local trade in natural products, livelihoods and poverty alleviation in a semi-arid region of South Africa. *World Development*, 36(3), 505–526. <https://doi.org/10.1016/j.worlddev.2007.03.003>
- Sheridan, M. J., & Nyamweru, C. (eds) (2008). *African Sacred Groves: Ecological Dynamics and Social Change*. Athens, USA: Ohio University Press.
- Steffen, W., Richardson, K., Rockstrom, J., Cornell, S. E., Fetzer, I., Bennett, E. M., ... Sorlin, S. (2015). Planetary boundaries: Guiding human development on a changing planet. *Science*, 347, 1259855. <https://doi.org/10.1126/science.1259855>
- Stenseke, M. (2016). The Intergovernmental science-policy platform on biodiversity and ecosystem services and the challenge of integrating social sciences and humanities. *Bulletin of Geography. Socio-Economic Series*, 33(33), 119–129.
- Sterling, E. J., Filardi, C., Toomey, A., Sigouin, A., Betley, E., Gazit, N., ... Jupiter, S. D. (2017a). Biocultural approaches to well-being and sustainability indicators across scales. *Nature Ecology and Evolution*, 1(12), 1798–1806. <https://doi.org/10.1038/s41559-017-0349-6>
- Sterling, E. J., Betley, E., Sigouin, A., Gomez, A., Toomey, A., Cullman, G., ... Porzecanski, A. L. (2017b). Assessing the evidence for stakeholder engagement in biodiversity conservation. *Biological Conservation*, 209, 159–171. <https://doi.org/10.1016/j.biocon.2017.02.008>
- Stern, P. C., Kalof, L., Dietz, T., & Guagnano, G. A. (1995). Values, beliefs, and proenvironmental action: attitude formation toward emergent attitude objects. *Journal of Applied Social Psychology*, 25(18), 1611–1636. <https://doi.org/10.1111/j.1559-1816.1995.tb02636.x>
- Suich, H., Howe, C., & Mace, G. (2015). Ecosystem services and poverty alleviation: A review of the empirical links. *Ecosystem Services*, 12, 137–147. <https://doi.org/10.1016/j.ecoser.2015.02.005>
- Summers, J. K., Smith, L. M., Case, J. L., & Linthurst, R. A. (2012). A review of the elements of human well-being with an emphasis on the contribution of ecosystem services. *Ambio*, 41(4), 327–340. <https://doi.org/10.1007/s13280-012-0256-7>
- Tanner, C. (1999). Constraints on environmental behaviour. *Journal of Environmental Psychology*, 19(2), 145–157. <https://doi.org/10.1006/JEVP.1999.0121>
- TEEB (2010). *The Economics of Ecosystems and Biodiversity: Ecological and Economic Foundations*. P. Kumar (ed.), London: Earthscan. Retrieved from <http://www.teebweb.org/publication/mainstreaming-the-economics-of-nature-a-synthesis-of-the-approach-conclusions-and-recommendations-of-teeb/>
- Tengö, M., Johansson, K., Rakotondraso, F., Lundberg, J., Andriamaherilala, J.-A., Rakotoarisoa, J.-A., & Elmquist, T. (2007). Taboos and forest governance: Informal protection of hot spot dry forest in Southern Madagascar.

- Ambio*, 36(8), 683–691. Retrieved from [http://www.bioone.org/doi/abs/10.1579/0044-7447\(2007\)36\[683:TAFGIP\]2.0.CO;2](http://www.bioone.org/doi/abs/10.1579/0044-7447(2007)36[683:TAFGIP]2.0.CO;2)
- UN (2015). *Transforming our World: the 2030 Agenda for Sustainable Development*. New York, USA: United Nations.
- Van Hecken, G., Bastiaensen, J., & Windey, C. (2015). Towards a power-sensitive and socially-informed analysis of payments for ecosystem services (PES): Addressing the gaps in the current debate. *Ecological Economics*, 120, 117–125. <https://doi.org/10.1016/j.ecolecon.2015.10.012>
- Verschuuren, B., Subramanian, S. M., & Hiemstra, W. (eds) (2014). *Community Well-being in Biocultural Landscapes*. Rugby, UK: Practical Action Publishing. <https://doi.org/10.3362/9781780448374>
- Vetter, S. (2013). Development and sustainable management of rangeland commons – aligning policy with the realities of South Africa’s rural landscape. *African Journal of Range & Forage Science*, 30(1–2), 1–9. <https://doi.org/10.2989/10220119.2012.750628>
- Vogl, A. L., Goldstein, J. H., Daily, G. C., Vira, B., Bremer, L., McDonald, R. L., ... Cassin, J. (2017). Mainstreaming investments in watershed services to enhance water security: Barriers and opportunities. *Environmental Science & Policy*, 75, 19–27. <https://doi.org/10.1016/j.envsci.2017.05.007>
- White, J., & White, B. (2012). Gendered experiences of dispossession: Oil palm expansion in a Dayak Hibun community in West Kalimantan. *Journal of Peasant Studies*, 39(3–4), 995–1016. <https://doi.org/10.1080/03066150.2012.676544>
- Whitmarsh, L. (2009). Behavioural responses to climate change: Asymmetry of intentions and impacts. *Journal of Environmental Psychology*, 29(1), 13–23. <https://doi.org/10.1016/j.jenvp.2008.05.003>